

Advanced Oxidation Processes (AOPs) for Refinery Wastewater Treatment Contains High Phenol Concentration

Alif Nurul Azizah¹ and I Nyoman Widiasta^{2*}

¹Chemical Engineering Department, Engineering Faculty, University of Diponegoro, Jln Prof. Soedarto, SH. Semarang, Central Java, Indonesia, 50239, Telp: (+62) 24 7460058

² Chemical Engineering Department, Engineering Faculty, University of Diponegoro, Jln Prof. Soedarto, SH. Semarang, Central Java, Indonesia, 50239, Telp: (+62) 24 7460058

Abstract. Petroleum Refinery wastewater is characterized by a high phenol content. Phenol is toxic and resistant to biological processes for treatment of the petroleum refinery wastewater. The combination of an AOP and a biological process can be used for treatment of the refinery wastewater. It is necessary to conduct a study to determine the appropriate condition of AOP to meet the phenol removal level. Two AOP configurations were investigated: H₂O₂ / UV and H₂O₂ / UV / O₃. From each process samples, COD, phenol and pH were measured. The oxidation was carried out until the targeted phenol concentration of treated effluent were obtained. The better result obtained by using process H₂O₂ / UV / O₃ with the H₂O₂ concentration 1000 ppm. After 120 minutes, the final target has been achieved in which phenol concentration of 37.5 mg/L or phenol degradation of 93.75%.

1 Introduction

Currently increasing of fuel consumption of about 5% per year make Indonesia have a very high fuel deficit if there is no additional oil processing capacity in the country. As a state-owned enterprise engaged in the energy sector, Pertamina continues to work hard in sustaining the availability of public fuel and achieving national energy security. Various oil and gas projects were established, one of which is the Residual Fluid Catalytic Cracking project, known as RFCC Project in Cilacap, Central Java. The RFCC refinery will increase premium production to 91,000 barrels per day or up 30,000 barrels per day (Pertamina.com).

The oil refinery consumes a minimum of 75.7 liters to 227.1 liters of clean water to process 1 barrel of crude oil [1]. The amount of waste water produced is about 41% of the amount of water used in the process [2,3]. From several previous laboratory tests it was found that refinery waste water had COD concentrations of 750-1600 mg / L, BOD about 300 mg / L, oil and mud concentrations 500-3000 mg / L, sulphide 13-17 mg / L, ammonia 14 mg / L, phenol around 960 mg / L, temperature around 30°C and pH above 7 [4-6].

The use of AOPs is proven to be effective for wastewater treatment, but it tends to be expensive because intermediate oxidation is formed during the process and makes a long chemical degradation, requiring considerable energy [7]. AOPs technology is

one or a combination of several processes such as ozone (O₃), hydrogen peroxide (H₂O₂), UV radiation, titaniumoxide (TiO₂-photocatalyst), Fenton process, as well as several other processes that can produce hydroxyl radicals). OH radical is the active species that has the highest relative oxidation potential, which is about 2.8 V, bigger than ozone having an oxidation potential of 2.07 V. This makes the OH very easy to react with other compounds around it [8].

The UV / O₃ / H₂O₂ method as an alternative technology and further development of UV / O₃ and UV / H₂O₂ methods. The AOPs technology was developed to obtain a rapid and nonselective concentration or amount of hydroxyl radicals for processing and removing persistent contaminants, as well as pollutants of EDCs (endocrine disrupting contaminants). Oxidizing agents, H₂O₂ and O₃ are added to the body of the liquid to be treated, and then activated with UV radiation photon energy so that spontaneously generate hydroxyl radical (•OH).

RFCC wastewater has a high phenol content. Phenol is toxic and resistant to biological processes, and required further process in order to be biologically processed in the existing refinery wastewater plant [9]. The use of biological treatment processes will be constrained by high levels of phenol which interfere the nitrification process. Some toxic compounds that can inhibit the nitrification process include: cyanide,

* Corresponding author: widiasta@che.undip.ac.id

thiourea, phenol, aniline, and heavy metals such as silver, mercury, nickel, chrome, copper and zinc.

In order to be biologically processed, wastewater must contain Dissolved Oxygen (DO) about above 2 mg / L, COD 500-3000 mg / L, ammonia 20 mg / L, phenol 50-100 mg / L, temperature around 30°C and pH about 5-10 [10-12]. In dealing with the problem proposed biological system of two stages. There was the addition of aeration tubs that require larger place and longer time [11].

The combination of AOPs processes and biological processes can be used in refinery wastewater treatment [13]. The main drawback of the AOPs process is the high cost due to large energy consumption. Selection of AOPs processes and operating conditions need to take into account the most extreme wastewater characteristics. From the existing wastewater data the process UV / H₂O₂ and UV / H₂O₂ / O₃ process have to be studied. The oxidation of organic pollutants in wastewater by AOP process utilizes the hydroxyl radical (OH) obtained by combining processes such as ozone O₃, UV, and H₂O₂ [14]. The number of OH radicals formed is influenced by the type of reaction and the presence of a strong oxidizer in the form of a mixture of hydrogen peroxide with ultraviolet light, a mixture of ozone with ultraviolet light and a mixture of ozone with hydrogen peroxide.

Phenol is classified in alcohol so can be oxidized to ketones, aldehydes and carboxylic acids [15]. Morkini et al., [16] states that carbon in ortho positions and the phenol compounds are electron donors so in such positions the electrons has high density. Hence the OH radical has a tendency to strike that position forming

intermediate compounds of catechol, hydroquinone, and benzokuinon. The end product of the oxidation reaction is carboxylic acid and ketone which can be processed in biological process [17]. The phenol degradation reaction is a first-order reaction, so the oxidizing concentration and reaction time will direct the fast-moving and precise reaction to form carboxylic acids and ketones. The more oxidizing concentrations and reaction time may not necessarily produce the desired product, whereas to obtain low phenol content requires high energy and cost. It is necessary to conduct a study to determine the condition of AOPs process that meets the target.

The experiment will evaluate the processing of AOPs with UV / H₂O₂ and UV / H₂O₂ / O₃ process and determine the operating conditions of AOPs target effluent with phenol concentration 50-100 mg / L.

2 Experimental

2.1 Materials

This experimental uses synthetic and real wastewater. Synthetic wastewater was made by dissolving phenol PA with deionized water. The real wastewater come from Pertamina RU IV Cilacap. Hydrogen peroxide (technical). UV Lamp Model Viqua Sterilight Silver S12Q-PA UV System 15 gpm Serial 131219515, 110-240v 50/60 Hz. Ozonator model Viqua Sterilight Model S8Q-OZ/2 Serial 13060530, 110-240v 50/60 Hz; Fe₂SO₄; H₂SO₄ and NaOH (*Analytical Grade*) as pH controller.

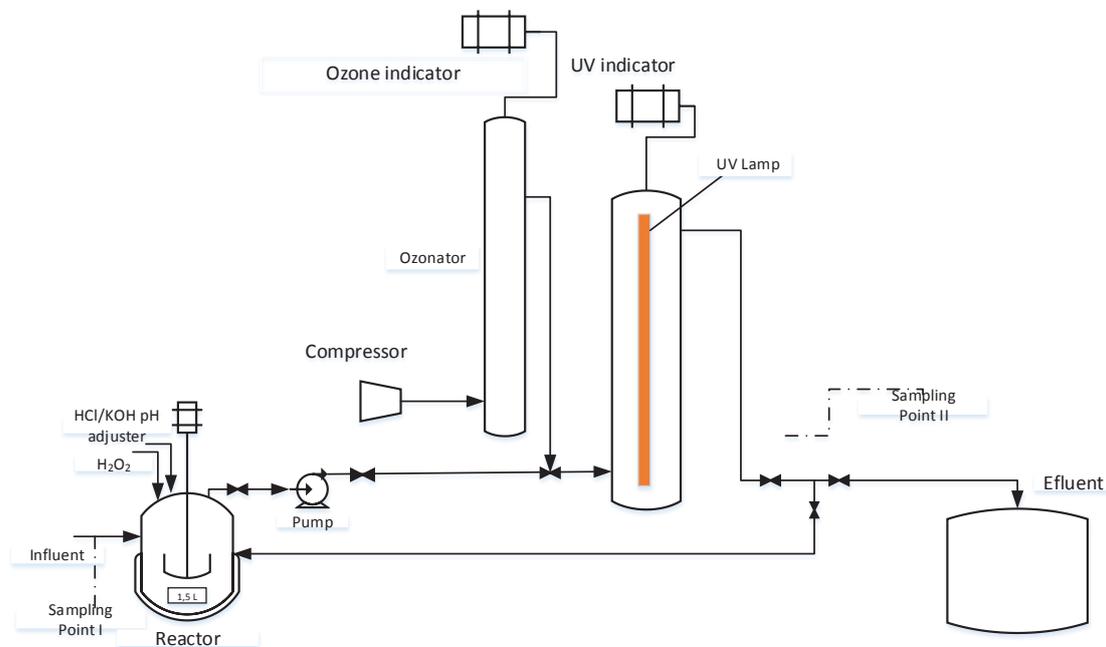


Fig 1. Advanced Oxidation Processes System

2.2 Method

Pertamina refinery waste water is first separated from solid impurities such as sand, stone, and solid waste. Prior to processing, wastewater was carried out by preliminary test, ie COD, phenol, and pH. A total of 20 liters of waste water goes into the reactor for AOPs processing using H₂O₂, ozone, and UV. Wastewater (influent) analyzed the COD, phenol, and pH content first. Then added H₂O₂ on various variables and arranged at pH using HCl and KOH. Influent is pumped into a UV process that is equipped with an ozone generator. (See Fig 1.).

After the process is done the analysis of effluent characteristics. The process is circulated with various time and hydrogen peroxide variables then analyzed effluent characteristics to obtain the required conditions to achieve the biological treatment on existing waste water treatment plant. The experiments were carried out with variations of the combination of H₂O₂ / UV and H₂O₂ / UV / ozone.

For Fenton system, prepare 250 ml wastewater on Schotts Bottle, then adjust pH into 3 by adding H₂SO₄. Add the Fe₂SO₄ catalyst then hydrogen peroxide for each variable. This experiment run with time variable from 30 until 180 minutes. The results analyzed as like as the AOPs process.

2.2 Analytical

Influent and Effluent are analyzed Phenol, COD, and pH using Hanna Instruments HI 3864, HI 93754C-25 HR, and HI 8424 in order.

3 Results and discussion

3.1 UV/H₂O₂ process

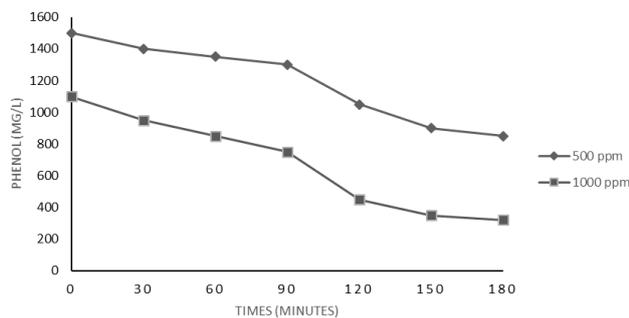


Fig 2. Phenol Degradation for synthetic RFCC wastewater by adding 500 mg/L H₂O₂ and 1000 mg/L H₂O₂

Fig 2. illustrates the experimental values of phenol final concentration in those wastewaters which have been put under advance oxidation process for reaction time 180 minutes. It can be concluded from Fig. that increasing the reaction time and the higher H₂O₂ concentration can lead more phenol degradation and reduction the final concentration of phenol in wastewater. The increase in H₂O₂ concentration

increased the phenol degradation and COD removal. Both 500 mg/L and 1000 mg/L H₂O₂ have not reach targetted phenol concentration yet. The initial phenol concentration 1100 mg/L decreases to 320 mg/L and COD removal of 2090 mg/L from initial COD 2186 mg/L.

H₂O₂/UV process was chosen for this purpose because it is known to be simple and effective for the removal of moderate to low amounts of nonbiodegradable organic compounds. Eliminate the ozon process means energy and cost saving, so it will be reliable if H₂O₂/UV process has similar result with H₂O₂/UV/O₃ process. Hence, it can be concluded that the H₂O₂/UV process needs longer time and higher H₂O₂ concentration removed the phenols from the solution to levels below the target.

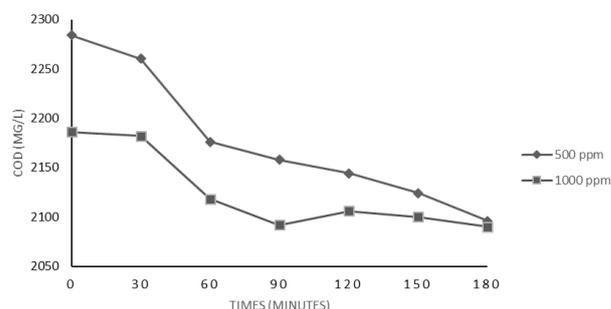


Fig 3. COD removal for synthetic RFCC wastewater by adding 500 mg/L H₂O₂ and 1000 mg/L H₂O₂

Figures 3. We can see that the COD removal show the descending trend against increasing the H₂O₂ concentration. By adding 500 mg/L H₂O₂ concentration just remove a little amount of COD. While adding more concentration to 1000 mg/L, COD removal reach 2090 mg/L in 180 minutes. From the figures above can be concluded that the COD removal show the descending trend against increasing the H₂O₂ concentration.

However, when H₂O₂ concentration exceeded the optimum concentration, amount of phenol degradation and COD removal was reduced. Although scavenging effect of hydrogen peroxide towards •OH occurs at higher concentrations [18], this is expected to give asymptotic values for phenol degradation. But reduction in phenol and COD removal observed may be due to the reaction of •OH and H₂O₂ and combination two •OH to form H₂O₂ [19]. In the other hand, the formation of •OOH which are significantly less reactive than hydroxyl radicals also reduces the reaction rate.

3.2 UV/H₂O₂/O₃ process

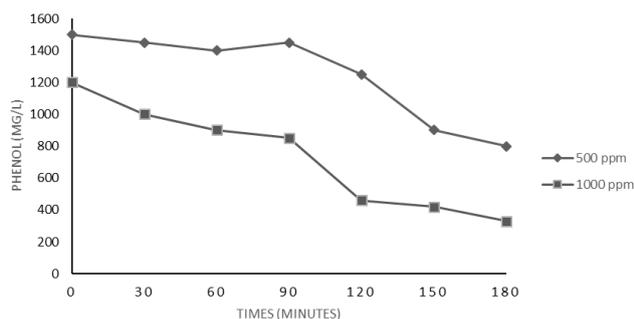


Fig 4. Phenol Degradation for synthetic RFCC wastewater by adding 500 mg/L H₂O₂ and 1000 mg/L H₂O₂

The effect of H₂O₂ concentration on the phenol degradation are depicted in Fig 4. It can be seen from the figures that the phenol concentration show the descending trend against increasing the H₂O₂ concentration. By adding 500 mg/L H₂O₂ we get phenol degrades from 1500 mg/L to 800 mg/L. Then by adding 1000 mg/L H₂O₂ we get phenol degrades from 1200 mg/L to 330 mg/L. In other hand COD removal on synthetic RFCC wastewater a bit lower too. COD decreases from 2360 mg/L to 2130 mg/L with 500 mg/L H₂O₂ and 2360 mg/L to 2072 mg/L by adding 500 and 1000 mg/L H₂O₂ respectively.

The phenol degradation using the H₂O₂/UV and H₂O₂/UV/O₃ are not having significant result. Phenol degradation using H₂O₂/UV process has efficiency 70,9 % then using H₂O₂/UV/O₃ process has efficiency 72,5 %. Irradiation of H₂O₂ with UV generated •OH radicals with powerful oxidizability, so the time for phenol degradation was higher those for H₂O₂/UV/O₃. The •OH radicals generated by photocatalysis were distributed more widely than the •OH radicals generated from dissolved ozone [20]. Ozone is unstable in solution, and •OH radicals are generated by an autolytic process [21].

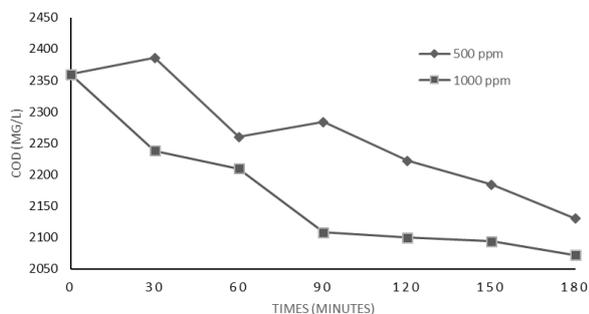
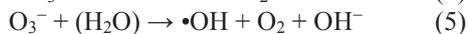
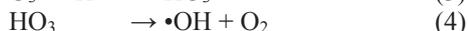
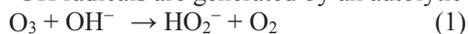


Fig 5. COD removal for synthetic RFCC by adding 500 mg/L H₂O₂ and 1000 mg/L H₂O₂

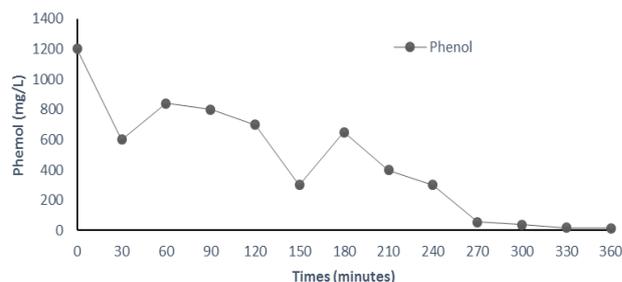
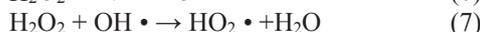
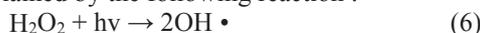


Fig 6. Phenol degradation for real RFCC by adding 500 mg/L H₂O₂

From the figure 6 we can see phenol degradation on real RFCC wastewater mostly symmetrical with the synthetic RFCC wastewater. Although we get a lower degradation there. The reason for this variation could be explained by the following reaction :



Theoretically, if the initial H₂O₂ concentration increases, higher hydroxyl radicals can be produced and phenol can be converted to CO₂ and H₂O more easily.

Fig 6. illustrates the experimental values of phenol concentration in RFCC wastewater which have been put under H₂O₂/UV/O₃ process for different reaction time. It can be conclude from Fig 6. that increasing reaction time will lead more phenol degradation and reducing the final concentration of phenol.

However there is an targeted condition of hydrogen peroxide concentration and reaction time. We see on Fig 6, adding 500 mg/L hydrogen peroxide have not reach target yet. The lowest phenol concentration reach at reaction time 150 minutes with 300 mg/L or degraded 57% from the initial concentration. While, by adding more 500 mg/L hydrogen peroxide on the next reaction reached to 55 mg/L along reaction time 270 minutes. The Ghaly et al. experimental observation confirmed the above statement. It can be clearly seen from Fig 2. that high initial concentration of H₂O₂ consumes the hydroxyl radicals according to desired reaction (7) and reduces the phenol degradation extent (increases the final concentration of phenol as depicted in Fig 6.)

The degradation mechanism of the real RFCC wastewater organic compounds, mostly aromatic-containing structures, is an electrophilic •OH addition to the rings that results in benzene ring conversion to phenol and then to dihydroxybenzene derivatives [22]. Both benzene and phenol produces dihydroxybenzene derivatives, along with quinines and several ring-opened products (namely carboxylic acids groups) [22;23]. This implies that as phenol is degraded, more is generated by the degradation of benzene. The direct reaction of •OH with the phenol molecules also gives rise to phenoxo radical. This radical, as detailed by Hosseini et al. (2007), is in resonance with radical structures in *o*- and *p*- positions [24] and becomes the starting point for the formation of different intermediates. More hydroxylated compounds such as catechol, hydroquinone, benzoquinone or hydroxyl hydroquinone are then

produced from subsequent reactions of the intermediates with $\bullet\text{OH}$ [25]. So, the degradation rate are less than the synthetic RFCC wastewater.

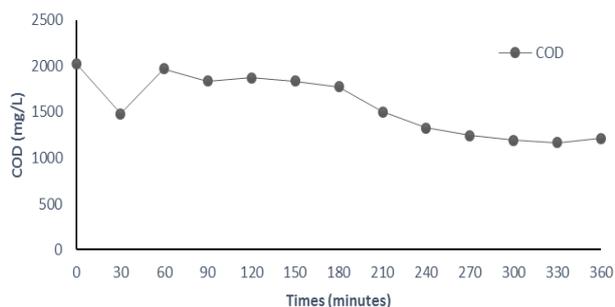


Fig 7. COD removal for real RFCC by adding 500 mg/L H₂O₂

Fig 7. illustrates COD removal as a function of time at different H₂O₂ concentration. COD removal are approximately symmetrical with the phenol degradation. The plots show that increasing reaction time will lead more COD removal and reducing the final concentration COD. By adding 500 mg/L H₂O₂ concentration just remove a little amount of COD typically same with the phenol degradation has not reach the target yet with this condition. While adding more concentration 500 mg/L, COD removal reach 1166 mg/L in 3300 minutes.

From the figures above can be seen from that the COD removal show the descending trend against increasing the H₂O₂ concentration. Removal efficiencies are reduced at 180 minutes reaction time. It may cause by the alkaline media that dissociated form of hydrogen peroxide (HO₂⁻) reacts with hydroxyl radicals (reaction 2.) more than two orders of magnituded faster than observed hydrogen peroxide, leading loss of carbon oxidation potential which is confirmed by a slight decrease in COD removal under alkaline condition [26].

4 Conclusion

Phenol degradation in RFCC wastewater using AOP was found to be possible. For UV/H₂O₂/O₃ process target phenol concentration 55 mg/L easily reached by adding 500 mg/L H₂O₂ along 180 minutes then adding more 500 mg/L H₂O₂ along reaction time 90 minutes. For H₂O₂/UV process both 500 mg/L and 1000 mg/L H₂O₂ have not reach targetted phenol concentration yet. The initial phenol concentration 1100 mg/L decreases to 320 mg/L and COD removal of 2090 mg/L from initial COD 2186 mg/L. It can be concluded that the H₂O₂/UV process needs longer time and higher H₂O₂ concentration removed the phenols from the solution to levels below the target. In other hand COD removal on synthetic RFCC wastewater a bit lower too. This implies that as phenol is degraded, more is generated by the degradation of benzene. The direct reaction of $\bullet\text{OH}$ with the phenol molecules also gives rise to phenoxy radical.

The research supported by PT. Pertamina RU IV Cilacap, is greatly acknowledge. The author acknowledge the laboratory

equipments provided by Chemical Engineering Departement of Diponegoro University Semarang that have resulted in this article.

References

1. C.H. Guernsey & Company, Water/Energy Sustainability Forum, (2009)
2. T. Doggett,, A. Rascoe, *Global Energy Demand Seen up 44 Percent by 2030*, (2009).
3. C. Marcilly, J. Catalyst. **216**, 47–62 (2003)
4. A. D. al deen A. P. Palaniandy, A. H. Bin Abdul, F. Shaik, J. Env. Chem. Eng., **3** 1117-1124, (2015)
5. A. M. Mansouri, F. Shahrezaei, A. A. L. Zinatizadeh, A. A. Hemati, M. Pirsaeheb, K. SharafiJ. Taiwan Inst. Chem. Eng. **45**; 2501-2510 (2014)
6. Z. Wasi, Khan, I. Najeeb, M.Tuiyebayeva, Z.Makhtayeva, Process Saf. Environ. Prot., (2014).
7. I. Muñoz, J. Rieradevall, F. Torrades, J. Peral, X . Doménech, Solar Energy,**79**, 369–75, (2005)
8. C. Gottschalk, JA. Libra, A. Saupe, *Ozonation of water and waste water: a practical guide to understanding ozone and its applications*, John Wiley & Sons, (2009)
9. M. Lapertot, C. Pulgarín, P. Fernández-Ibañez, MI. Maldonado, L. Pérez-Estrada, I Oller, *Water Res*; **40**:1086–94, (2006)
10. G. Bitton., *Wastewater Microbiology*. New York: A John Wiley & Sons, INC., (1994)
11. I. Oller, S. Malato , J.A. Sánchez-Pérez, Sci. Total Environ. **409**, 4141–4166, (2011)
12. Anonim, Data Desain WWTP-Ipal, Pertamina, (2017)
13. V. Sarria, M. Deront, P. Péringer, C. Pulgarin, Appl. Catal. B. Environ., **40**, 231–46, (2003)
14. W. H. Ghaze, J-W. Kang, D.H. Chapin, *The chemistry of water treatment processes involving ozone, hydrogen peroxide and ultraviolet radiation*, (1987)
15. Isyuniarto, W. Usada, Suryadi, A. Purwadi, Degradasi Fenol dalam Limbah Pengolahan Minyak Bumi dengan Ozon, P3TM BATAN Yogyakarta (2003).
16. A. Morkini, D. Oussi, E. Chamarro, S. Esplugas, Photoxidation of Phenol Uv Radiation Aqueous Solution, <http://www.photobiology.com/IUPAC98/esplugas/phenol.html>. (1998)
17. M. Y. Ghall, G. Hartel, R. Hasender, Photochemical oxidation od p-chlorophenol bu UV/H2O2 and photo-Fenton process. *A comparative study, Waste Management* (New York, N.Y.) **21** 41-7, 11150131, (2001)
18. B. Subramanian, Q. Yang, Q. Yang, A.P. Khodadoust, D.D. Dionysiou, J. Photochem. Photobiol. A: Chem. **192** 114–121, (2007)
19. X.R. Xu, X.Y. Li, X.Z. Li, H.B. Li, Degradation of melatonin by UV, UV/H2O2, Sep. Purif. Technol. **68** 261–266, (2009)
20. Harufumi S., Sadao A., Hideki Y., Journal of Water Process Engineering **7** 54–60, (2015)

21. Bai C., Xiong X., Gong W., Feng D., Xian M., Ge Z., Xu N., *Desalination* **278** 84–90 (2011).
22. Aranda, E., Marco-Urrea, E., Caminal, G., Arias, M.E., García-Romera, I., Guillén, F., *J. Hazard. Mater.* **181**, 181–186, (2010)
23. Z. Zhang, Q. Xiang, H. Glatt, K. L. Platt, B. D. Goldstein, G. Witz., *Free Radic. Biol. Med.*, **18**, 411–419, (1995)
24. S. N. Hosseini, S. M. Borghei, M. Vossoughi, N. Taghavinia, *Appl. Catal. B: Environ.*, **74**, 53–62, (2007)
25. A. M. Peiró, C. Domingo, J. PeralX. Domènech, E. Vigil, M.A. Hernández-Fenollosa, M. Mollar, B. Mari, J.A. Ayllón, *Thin Solid Films* **483**, 79–83, (2005)
26. K. Kosala, H. Yamada, K. Shisida, S. Echigo, R.A. Minear, H. Tsuno, M. Saburo, *Water Res.*, **35** 3587, (2001)